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NanoEHS beyond toxicity – focusing on biocorona

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The first phase of studies on environmental health and safety of nanomaterials (nanoEHS) has been mainly focused on evidence-based investigations that probe the impact of nanoparticles, nanomaterials and nano-enabled products on biological and ecological systems. The integration of multiple disciplines, including colloidal science, nanomaterials science, chemistry, toxicology/immunology and environmental science, is necessary to understand the implications of nanotechnology for both human health and the environment. While strides have been made in connecting the physicochemical properties of nanomaterials with their hazard potential in tiered models, fundamental understanding of nano-biomolecular interactions and their implications for nanoEHS is largely absent from the literature. Research on nano-biomolecular interactions within the context of natural systems not only provides important clues for deciphering nanotoxicity and nanoparticle-induced pathology, but also presents vast new opportunities for screening beneficial material properties and designing greener products from the bottom up. This review highlights new opportunities concerning nano-biomolecular interactions beyond the scope of toxicity.

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Environmental significance

The current literature concerning environmental health and safety of nanomaterials (nanoEHS) is often focused on describing the phenomenological consequences of nanomaterial exposure and the toxicity of nanoparticles to cells and model organisms. A wealth of information on nano-biomolecular interactions has not been fully exploited making it difficult to identify nano-specific effects as well as trace the origin of toxicity. We believe that continued development of nanoEHS hinges on a critical extension from reporting macroscopic and microscopic phenomena to understanding nano-biomolecular interactions. This new frontier will facilitate resolving the interplay between nanomaterials and natural systems that elicits beneficial or detrimental effects on human and environmental health.

NanoEHS, a reflection

The birth of the nanoEHS (Environmental Health and Safety of Nanomaterials)¹ field was originated by the concern on the unknown consequences of introducing engineered nanomaterials (ENMs) to the human body or the environment.^{2–5}

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^b Bren School of Environmental Science and Management, Earth Research Institute and University of California Center for the Environmental Implications of Nanotechnology (UC CEIN), University of California, Santa Barbara, California 93106, USA It is a natural question to ask whether the very properties of ENMs being exploited in nano-related applications (such as high surface reactivity and ability of crossing biological barriers) might also have negative health and environmental impacts.² In order to answer this question, the field took an initiative to develop assessment tools based on traditional chemical toxicology and was set to investigate the impacts of ENMs and nanotechnology from cradle to grave at a pace commensurate with the development of nanotechnology.^{6,7} It was guickly recognized that the impact of ENMs on biological systems, positive or negative, hinges on biophysicochemical interactions at the nano-bio interface.8 In contrast to traditional chemicals, ENMs possess diverse physicochemical properties that require additional consideration in toxicity testing. The dynamic physicochemical properties of ENMs play a critical role in their fate and transport, human and environmental exposure conditions and hazard generation.^{1,9} All of these aspects require extensive knowledge acquisition about the unique interactions between ENMs and various biological systems. Therefore, by nature, the path set forth for this field requires an integration of disciplines, including



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colloidal science, materials science, chemistry, biophysics, toxicology/immunology, and environmental science.^{1,7,10,11}

Because of the interest in discovering and the economic incentive of avoiding the hazardous effects of ENMs, this multidisciplinary enterprise of nanoEHS has been focused on toxicity assessment. A decade later, with a wealth of information accumulated on various types of ENMs and their hazard profiles obtained using tiered models, *i.e.* from cells to animals, we now reflect on the advances of this field, summarize the effective research approaches, and anticipate new research topics going forward. We envisage that the knowledge gathered and being pursued at the nano-bio interface will have the potential to aid in inventing new tools for treating human diseases as well as preventing environmental crises.

Predictive approach and toxicological paradigm

With the "nano" component of the nano-bio interface being defined as anything man-made and at least one dimension falling in the range of 1–100 nm, the "bio" component expands from proteins and cells to low-level organisms and vertebrates, and up to mammals and humans (Fig. 1). Considering the vast number of nano-bio interactions that could be generated when ENMs with different physicochemical properties such as composition, size, surface area, shape, aggregation/agglomeration, crystallinity and surface functionalization make contact with various biological entities, such as lipids, proteins, cell membranes, organelles, tissues and whole organisms, it is not possible to describe every aspect of nanobio interactions.⁸ Knowing the conundrum of the chemical industry, where among more than half a million industrial chemicals invented and manufactured, fewer than a thou-

sand of them have undergone toxicity testing, it was rational to invest in developing alternative testing strategies for ENMs. In agreement with the landmark 2007 report from the US National Academy of Sciences, "Toxicity Testing in the 21st Century: A Vision and a Strategy", tiered testing strategies based on predictive toxicological approaches were implemented to examine nano-bio interactions at incremental levels, *i.e.* from cells to organisms.¹² The rationale was that by using *in vitro* platforms (high-throughput screening, HTS), a variety of nano-bio interactions can be studied simultaneously in order to prioritize more detailed *in vivo* investigations.^{13,14}

To make such predictive approaches valid and vital, translation and validation between in vitro and in vivo testing is essential. Along this direction, a few representative toxicological paradigms have emerged over the years.^{10,13} For instance, in vitro assays have revealed the crucial role of the release of ions in the toxicity of certain metal and metal oxide-based ENMs, e.g., Ag, CuO and ZnO NPs.¹⁵⁻²² The release of metal ions from NPs at the nano-bio interface has been often linked to generation of reactive oxygen species (ROS) which set forth a series of cellular responses leading to oxidative injury and cell death.^{23,15-20} These responses have also been observed in vivo, proving that in vitro assays can successfully predict toxicological pathways.²⁴⁻²⁸ For insoluble ENMs, bandgap structures and overlap of the conduction bands of ENMs with biological redox potential determine the generation of detrimental ROS.²⁹ Among the surface properties of ENMs, surface charge and functionalization (e.g., PEGylated vs. bare particles) serve as main contributors to nano-bio interactions and subsequent toxicological outcomes.30 Membrane damage caused in in vitro experiments by cationic



Fig. 1 A new frontier of research on nanoEHS. Much fundamental science and their implications remain to be understood and exploited at the interface between ENMs and the biocorona, a dynamic layer of biomolecules adsorbed on the NP surface through self-assembly, crowding and, at times, aggregation and fibrillation. It is important to understand that the passage and fate of NPs inside cells and tissues and across the barriers between different compartments, organs as well as species are consequential to the physicochemical and biological identities of the ever-evolving biocorona.

ENMs such as polystyrene NPs, liposomes, dendrimers, mesoporous silica, and carbon nanomaterials has been shown to translate to *in vivo* toxicological outcomes.^{31–36} Evidently, a comprehensive grasp of biological injury mechanisms originating from the nano-bio interface was the key for developing such predictive approaches. It is important to note that the aforementioned toxicological paradigms were mostly derived from empirical data and known toxicological pathways. With the advancements in proteomics and genomics-based approaches, it is likely that new toxicological paradigms will soon become available.^{37–40}

High-throughput screening and the big data era

In order to provide discovery platforms with the ability to consider a wide range of nano-bio interactions, streamlining experimental procedures to reach HTS capacity has undergone development concurrently with the nanoEHS field. The benefit to implement HTS for nanoEHS research was consistent with the advocate for increased efficiency of toxicity testing by transitioning from qualitative, descriptive animal testing to quantitative, mechanistic, and pathway-based toxicity testing in cell lines.¹²

One in vitro HTS example that emerged was the study conducted by George et al. on a series of metal and metal oxide NPs.41 The in vitro HTS platform was established based on a cocktail of compatible fluorescent dyes to develop an epifluorescence-based approach that screened for a functionally connected series of oxidative stress-related injury responses in mammalian cells. These fluorescent dyes were selected to avoid possible interactions between the NPs and the assay reagents that might yield false positives and negatives.42 This multi-parameter HTS assay, which was composed of robotic handling of multitier plates, automated liquid handing and series of ENM dilutions, contemporaneously assessed the total cell number and nuclear size, mitochondrial membrane potential, intracellular calcium flux, and increased membrane permeability based on microscopy images captured automatically. The functional link between the cellular responses measured using the cocktail of fluorescent dyes makes such a platform suitable for assessing the biological effects of any ENMs with the potential of generating oxidative injury. The high-volume, image-rich, kinetic HTS data sets led to further development of machine-learning and hazard-ranking tools that are capable of establishing specific structure-activity relationships (SARs) of ENMs.43-46 Details on the statistical modelling will be discussed in a later section. Similarly, HTS platforms employing robotics, automated imaging and computer-assisted image analysis have also been established for algae, sea urchin and zebrafish embryos.⁴⁷⁻⁵⁰ It is also plausible that new HTS platforms for nanoEHS could be discovered based on proteomics and genomics-based approaches.51

Naturally, HTS has facilitated a smooth transition for the nanoEHS field to enter the big data era. Exploiting computational methods for assisting the establishment of quantitative structure–activity relationships (QSARs) and further enabling safer-by-design or risk reduction approaches will be discussed in the statistical modelling section.

The less explored: nano-bio interaction at the molecular level

With the aforementioned research approaches, the nanoEHS field has made significant advances illustrating interactions at various nano-bio interfaces. While majority efforts have emphasized either the "nano" component of the nano-bio interactions (i.e. establishing relationships between the "nano" properties and the biological outcomes) or the biological consequences at the cellular level and beyond (*i.e.* cytotoxicity, developmental disturbance, pathological outcomes in tissue culture and animal models), fundamental understanding of molecular-level nano-bio interactions, known generically as the protein corona or biocorona, remains less explored within the context of nanoEHS. While making connections between in vitro data and in vivo investigations, it is important to realize the differences between the simpler biocorona formed in in vitro assays and the much more complex protein-rich environment in *in vivo* testing.⁵² Moreover, since most ENM-cell interactions occur in a biomolecule-rich environment, it is important to recognize the formation of biocorona as inevitable simply for the argument of energetics. Understanding the identity and formation kinetics of biocorona in association with the physicochemical properties of ENMs not only is the key to deciphering the molecular mechanisms of NP-induced adverse effects but also presents vast opportunities to exploit the biocorona.^{53,54} In the following sections, we review research on the role of NP-biomolecular interactions in nanomaterial toxicity and discuss experimental, statistical modelling and computer simulation strategies capable of revealing the inner workings and nanoEHS implications of the biocorona. An outlook of this new frontier is presented at the end of the review.

NP-biomolecular interactions in nanotoxicology, a frontier

The importance of defining the key properties that drive the toxicity of ENMs has long been emphasized in both human and environmental toxicology.^{8,13} The central role of the biocorona – a dynamic layer of biomolecules adsorbed on the NP surface – in cell interactions and subsequent biological effects has been recognized in biomedical applications of ENMs, as highlighted by several excellent reviews.^{55–57} Recently, understanding the interplay between biomolecules and ENMs in environmental toxicity testing has been brought into focus.⁵⁸

NP toxicity influenced by laboratory test media

The role of the chemical composition of exposure media in the toxicity of ENMs has been acknowledged from early on. In 2006, Zhu *et al.*⁵⁹ reported a stimulatory effect on the growth of protozoan *Tetrahymena pyriformis* when cultured

with multiwall carbon nanotubes (MWCNTs) in proteose peptone yeast extract medium. This effect was attributed to the noncovalent binding of peptone with MWCNTs that facilitated cell entry of peptone by phagocytosis. In contrast, growth inhibition was observed when the protozoa were exposed to MWCNTs in filtered pond water. Other authors have also reported that high ionic strength or organic-rich media influence the agglomeration pattern of NPs, through modifying the surface chemistry of ENMs that often results in reduction of NP number and concentration and, consequently, a decrease of NP-cell contact and NP-specific toxicity.^{60,61} In the case of soluble metal-based NPs (e.g. CuO, ZnO, Ag, Fe₃O₄ and CdSe), medium composition and biocorona formation strongly influence the extent of dissolution, speciation and bioavailability of dissolved metal ions. Kakinen et al.62 compared 17 different test media for CuO NP speciation and bioavailability and concluded that dispersion of CuO NPs in the medium containing high concentrations of organic compounds resulted in enhanced dissolution of copper, due to increased dispersion of CuO NPs and complexation of copper ions with the organic compounds. These simultaneous processes influenced the level of bioavailable copper and consequently nanotoxicity, compared to that in media with little to no organic compound content. Thus, biocorona formation by organic compounds is an important factor that can influence the toxicity of NPs and should be taken into consideration when interpreting the results of toxicity tests.

NP toxicity influenced by natural organic matter (NOM)

To assess and predict NP behaviour in natural waters and soils, instead of conventional laboratory test media, experiments are conducted under more environmentally relevant conditions. Such test designs often include NOM which is a complex mixture of organic compounds originating from the decomposition of plant and animal residues, soluble microbial products, and extracellular polymeric substances (EPS).^{63,64} NOM is composed of nonhumic (polysaccharides, lignin, proteins, and polypeptides) and humic substances (high-molecular-weight microbial degradation products of plant material). Since the major acidic functional groups in NOM, carboxyl and phenolic groups, deprotonate at pH > 3, NOM is negatively charged in soils and natural waters. Due to the high adsorption affinity of NOM for mineral surfaces and π - π stacking on hydrophobic ENMs, coupled with its environmental relevance, NOM is often used as a dispersing and stabilizing agent of NPs in the aqueous phase.⁶⁵⁻⁶⁷ Furthermore, due to increased dispersibility or alternation in the polarity of ENMs, NOM coating has been shown to modify NP toxicity. For instance, humic acid mitigated the toxicity of MWCNTs to unicellular green alga Chlorella pyrenoidosa⁶⁸ and decreased the uptake of fullerene (nC_{60}) in *D. magna* and zebrafish.⁶⁹ In other studies, NOM-coated MWCNTs proved more toxic to daphnids⁷⁰ and a freshwater diatom⁷¹ than their pristine counterparts. NOM coating on the NP surface has been shown to exert cell type-specific uptake and damage: NOM-coated fullerene more readily entered mammalian cells in comparison to plant cells.⁶⁵ The latter was explained by the hydrophobic nature of NOM-coated NPs that facilitated partitioning into mammalian cell membranes but were excluded by plant cell walls. However, NOM does not solely govern the uptake of ENMs to cells and organisms: in a study with rice plants, NOM-coated C70 was taken up by plants and translocated through plant tissues while NOM-covered MWCNTs were merely adsorbed to the plant roots.⁶⁶ In the case of spherical metal NPs with well-defined size and shape, it is well known that surface charge is one of the main determinants of NP uptake and translocation in plants with positively charged NPs most readily taken up by plant roots.⁷² Since NOM coating renders NPs negatively charged, the reduced toxicity and accumulation of NOM-coated NPs in plants are often attributed to hindered interactions between NPs and negatively charged plant root surfaces.⁷³⁻⁷⁶ While in the latter studies NP exposure was conducted under defined hydroponic conditions, others have explored NP effects on plants and rhizospheres in more complex mesocosms consisting of NP-amended soil and organisms. Ge et al.77 showed that the plant root system significantly influenced NP effects on soil bacterial communities. The study revealed the complex nature of NP interactions in planted versus unplanted soil systems: plant growth ameliorated ZnO NP toxicity but promoted the impact of CeO2 NPs on bacterial communities, possibly by changing the belowground biogeochemistry. Living organisms are known to secrete extracellular compounds during their life cycle to facilitate nutrient uptake, intercellular communication, cell migration and other functions. EPS produced by microorganisms⁷⁸ and composed mainly of polysaccharides and proteins constitute an important fraction of NOM and thus have been studied for their effects on NP stability and toxicity. Compared to total NOM, EPS have been shown to improve the stability of NPs to a greater extent, including reduced aggregation of single-wall carbon nanotubes (SWCNTs)79 and copper-based NPs.80 Improved dispersion of EPS over NOM may be a result of preferential adsorption of protein-like components over humic substances, as demonstrated for graphene oxide and reduced graphene oxide.⁸¹ Increased residence times of EPS-dispersed ENMs in the water column could pose a higher risk to pelagic organisms through extended NP exposure. EPS can also mediate increased dissolution of metal-based NPs or leaching of metal catalysts from carbon nanomaterials to impact bioavailability.^{79,80} Other biomolecules that have been shown to disperse and disaggregate NPs through biocorona formation include alginate,82,83 bovine serum albumin (BSA)84 and monosaccharides.85

The physicochemical characteristics of NPs in environmental matrices can also be influenced by direct particle-cell contacts that may result in NP-microorganism heteroagglomerate formation or changed dispersion patterns of NPs. *Pseudomonas aeruginosa*, a Gram-negative bacterium, was shown to mediate the dispersion of TiO₂ NP homoagglomerates by preferential adsorption of the NPs to bacterial cells, highlighting the role of bacteria in enhancing NP mobility in the environment.⁸⁶ In contrast, *D. magna* reduced the stability of carbon nanotubes in freshwater, likely by digesting the nanotube biocorona.⁸⁷ Biomodification of NPs by both NOM and direct cell or organism contact is an important factor that influences the environmental fate and toxicity of ENMs.

Mitigation of NP toxicity by biocorona/extracellular substances

Biocorona formation generally renders NPs stably dispersed in the aqueous phase, potentially increasing NP contacts with cells and organisms that has been shown to enhance nanotoxicity.^{88,89} However, EPS produced by cells can act as a potent mitigator of nanotoxicity as reported by many studies. NOM-dispersed MWCNTs were shown to have strong affinity for EPS secreted by freshwater diatom Nitzschia palea, which was proposed as a protective mechanism against MWCNT toxicity.⁷¹ Similarly, the role of EPS in mitigating TiO₂ NP toxicity has been reported in cyanobacteria - the EPS-depleted mutant strain was more susceptible to NPs than the wild type.⁹⁰ Chemically inert TiO₂ NPs can become highly reactive under ultraviolet light and exert toxicity through ROS generation. Since ROS are generally short-lived, NPs have to be in close vicinity to the cell to exert ROS-mediated toxicity. This is likely the reason why extracellular compounds secreted by cells act as a protective layer by consuming harmful ROS.⁹⁰ The role of extracellular substances in regulating the biological effects of soluble metal NPs has also been extensively studied. Chen et al.91 showed that cell damage induced by ZnO NPs to freshwater algae Chlorella sp. was counteracted by algal exudates that coated positively charged NPs through electrostatic interactions and hydrogen bonding as well as complexed harmful Zn²⁺ ions. In addition, ZnO NPs were extensively agglomerated in the presence of algae, resulting in reduced dissolution of NPs. Similar mechanisms for mitigation of AgNP toxicity by algal extracellular dissolved organic carbon (DOC) compounds were proposed.92 Namely, freshwater alga Chlamydomonas reinhardtii was considerably less sensitive to AgNPs in the late growth phases when extracellular DOC levels were high, compared to cells in the early growth phases. It was proposed that algal exudates reduced AgNP toxicity by forming AgNP agglomerates, complexing toxic Ag⁺ ions, limiting particle-cell interactions and acting as a sink for NP-generated ROS. The protective role of EPS against AgNP toxicity also in bacterial biofilms was recently reported.93 In the marine diatom Thalassiosira weissflogii, the toxicity of AgNPs was caused solely by Ag⁺ ions, whereas the toxicity was mitigated by increased EPS production in response to Ag⁺ ions and subsequent complexation of toxic metal ions and ROS by EPS.94 An increase in EPS production has also been reported in freshwater alga Chlorella pyrenoidosa upon exposure to CuO NPs that generated ROS and caused mitochondrial depolarization in the alga through Cu²⁺ ions, indicating a possible cellular protective mechanism of EPS.95 The influence of EPS on the physicochemical

properties of CeO₂ NPs and AgNPs was recently studied by Kroll et al.,⁹⁶ who used EPS extracted from periphyton with varying composition to study changes in particle size and solubility upon incubation in EPS; EPS were found to stabilize CeO2 NPs and reduce their solubility. However, AgNPs increased in size in EPS, both at the single-particle level and through agglomeration. The increase in particle size was counterintuitively accompanied by higher dissolution of AgNPs. It was concluded that simultaneous oxidation of AgNPs and reduction of Ag⁺ occurred in the system, whereas EPS components, specifically polysaccharides with their hydroxyl and hemiacetal groups, contributed to the reduction of Ag⁺ to Ag⁰ at the surface of AgNPs that increased in size. The role of hemiacetal groups of sugars in reducing Ag⁺ in bacterial EPS was also reported by Kang et al.97 Reduction of Ag⁺ to less potent AgNPs by Escherichia coli EPS was interpreted as a mechanism for mitigating the antibacterial activity of silver. Such detoxification strategies are not confined to prokaryotic microorganisms. Juganson et al.98 showed that the extracellular soluble fraction of protozoa Tetrahymena thermophila contained potent Ag⁺-reducing substances that contributed to the formation of AgNPs - a less toxic form of silver to protozoa - from AgNO₃. While the carbohydrates of EPS might be responsible for the reduction of Ag⁺ to metallic silver, soluble proteins were shown to play a role in the formation and stabilization of AgNPs by rendering a protein corona on the NPs. The contributions of both proteins and carbohydrates to the formation of biogenic NPs were recently confirmed also in the case of selenium NPs synthesized in the EPS of anaerobic granular sludge.99 The functional groups of the EPS were shown to govern the surface charge of the NPs, thus influencing their fate in the environment. AgNPs readily interacted with proteins in the biological media resulting in the formation of a protein corona on the NP surface. These interactions may interfere with the activity of catalytic proteins, most likely through the release of Ag⁺ ions that reacted with the cysteine residues and N-groups of the enzyme to impair the enzymatic activity.¹⁰⁰ On the other hand, formation of protein corona has been demonstrated to reduce nanotoxicity to certain human cells. For instance, binding of a blood protein - bovine fibrinogen - rendered SWCNTs less harmful to a human acute monocytic leukemia cell line and human umbilical vein endothelial cells.¹⁰¹ Considering protein corona, and biocorona in general, there is a growing body of literature covering the role of NP biocorona in inducing physiological, including immunological, responses in mammalian organisms.^{102,103} However, studies that report organismal defence mechanisms to NOM-, EPSor biomacromolecule-coated NPs in environmentally relevant organisms are rare. Recently, NP-protein corona formation was characterized in the hemolymph serum of the marine bivalve Mytilus galloprovincialis.104 The study indicated that protein-covered polystyrene NPs induced higher levels of cell damage and ROS production than pristine polystyrene NPs. The higher toxicity of NPs with biocorona was attributed to dysregulation of immune responses. On the other hand,

another study in plants found that tannic acid coating reduced the toxicity of neodymium oxide NPs to pumpkin by boosting superoxide dismutase and catalase activity to harness excessive ROS.⁷³ Overall, there is a knowledge gap on the mechanistic understanding of how biocorona formation influences NP physiological effects in environmentally relevant species.

NP effects on the release and structure of specific extracellular substances

At sublethal concentrations, ENMs are known to modulate the expression and synthesis of specific compounds secreted by microorganisms. For instance, ZnO NPs were shown to inhibit biofilm formation and production of pyocyanine, a virulence factor, in P. aeruginosa.¹⁰⁵ Gene expression analysis revealed upregulation of genes involved in zinc cation efflux pump functioning and repression of pyocyanine-related operons. In an earlier study, Dimkpa et al.¹⁰⁶ showed that in a soil-beneficial bacterium P. chlororaphis ZnO NPs inhibited the synthesis of another virulence factor, pyoverdine, mainly through dissolved zinc ions. In contrast, CuO NP-induced inhibition of the same biomolecule synthesis was found to be nano-specific.^{106,107} In addition to being a virulence factor, pyoverdine functions as an iron-chelating compound that numerous bacteria secrete to acquire iron from the surrounding medium. Interestingly, according to Avellan et al.,¹⁰⁸ bacterial siderophores were also capable of chelating iron ions from iron-doped germanium-imogolite nanotube structures while the secreted biomolecules promoted the biodegradation of the nanotubes. This study suggests that metabolic responses to certain components of an ENM may also induce biodegradation of the ENM. Besides metal-based ENMs, sublethal concentrations of SWCNTs have been shown to inhibit the production of pyoverdine by P. aeruginosa,¹⁰⁹ at both physiological and gene transcriptional levels. These studies underline the impacts of ENMs on microbial physiology and cellcell interactions through EPS synthesis. Yet, molecular information concerning the composition, conformation, evolution and implications of such NP-biomolecule interactions remains to be elucidated.

Nanoparticle-biocorona – experimental strategies

Major techniques

A host of biophysical and analytical methodologies are available for quantifying and determining protein species adsorbed on NPs, for nanotoxicology, nanomedicine and nanobiotechnology applications.^{110,111} Despite differences in subject and - potentially - dosage, many of these methodologies and protocols may be utilized for examining biocoronae secreted by environmental organisms upon ENM exposure (Table 1). A survey of the literature shows that UV-vis spectrophotometry, polyacrylamide gel electrophoresis (PAGE) and sodium dodecyl sulfate-PAGE (SDS-PAGE), size exclusion chromatography (SEC), differential centrifugal sedimentation (DCS) and the highly sensitive proteomic mass spectrometry (MS) are some commonly used techniques for separating protein species among themselves and from NPs and for identifying and quantifying protein compositions within a corona.^{89,112-121} For imaging, transmission electron microscopy (TEM; resolution: 0.1-0.5 nm), atomic force microscopy (AFM; vertical resolution: 0.1 nm) (Fig. 2A and B) and confocal fluorescence microscopy (resolution: ~350 nm) can reveal information of soft and hard coronae on NP surfaces in either the dehydrated or aqueous phase, and often down to the single NP level for TEM and AFM.¹²²⁻¹²⁶ Dynamic light scattering (DLS; detection range: 0.3 nm-10 µm) and synchrotron-based techniques such as small-angle X-ray scattering (SAXS), small-angle neutron scattering (SANS) (Fig. 2C), and wide-angle X-ray scattering (WAXS) can yield information about the hydrodynamic size, polydispersity index, structure, morphology and crystallinity of NPs and polymeric materials (including proteins) on the mesoscopic scale.^{110,125,127} Shifts in frequency and amplitude of surface plasmon resonance (SPR; affinity range: fM to mM) may render association and dissociation constants of protein-NP binding where the NPs are immobilized on a gold-coated substrate.¹¹² Isothermal titration calorimetry (ITC) quantifies binding affinity, stoichiometry, and thermodynamic parameters of proteins interacting with NPs.^{113,122} Hyperspectral imaging, a dark-field microscopy technique, is suited for semi-

Methodologies			
Isotherm, affinity, stoichiometry,		Morphology, crystallinity,	
thermodynamic properties		dynamics	
UV-vis spectrophotometry	Transmission electron microscopy	Circular dichroism	Polyacrylamide gel electrophoresis
Dynamic light scattering	Scanning electron microscopy	Nuclear magnetic resonance	Size exclusion chromatography
Surface plasmon resonance	Atomic force microscopy	Small-angle X-ray scattering	Mass spectrometry
Isothermal titration calorimetry	Confocal fluorescence microscopy	Small-angle neutron scattering	Differential centrifugal sedimentation
Fourier transform infrared spectroscopy	Hyperspectral imaging	Wide-angle X-ray scattering	X-ray photoelectron spectroscopy
Thermogravimetric analysis Differential scanning fluorometry		X-ray crystallography	Raman spectroscopy

Table 1 List of major experimental methodologies for the study of biocorona



Fig. 2 Exemplified characterization of protein aggregation and NP-biocorona interactions. A: TEM imaging of AgNP-luciferase corona.¹⁰⁰ B: AFM phase imaging of ILQINS helical ribbons. C: SAXS spectra showing the formation of microcrystals converted from the ILQINS helical ribbons.¹²⁵ D: ICP-MS quantification of Zn ion release upon ZnO-algae interactions. Inset: SEM image of algal cells adsorbed with aggregates of ZnO NPs.⁹¹ E: Quantifying and analysing the transformation of NPs in a cellular environment using hyperspectral imaging (HSI); SL – spectral library; SAM – spectral angle mapper.¹²⁹ F: High-throughput DLS measuring the aggregation of human islet polypeptide over time.¹²⁶ G: Principle of DSF, where T_m represents the melting temperature of the protein and ΔT denotes the temperature range of the protein undergoing a transition from the molten to aggregated state.¹³⁷ H: ATR-FTIR spectra showing the binding of ZnO NPs and algal exudates.⁹¹

quantification of protein adsorption onto metal NPs, with or without the presence of cells or organisms.^{128,129} Fourier transform infrared spectroscopy (FTIR) measures the absorbance of multi-wavelength light to infer interactions and bond formation between proteins and NPs.⁹¹ Circular dichroism (CD) spectroscopy resolves changes to the secondary structure of proteins and peptides upon binding to NPs.^{130,131} Nuclear magnetic resonance (NMR) spectroscopy (for proteins <70 kDa) and X-ray crystallography (resolution: 1-5 Å, for small to arbitrarily large proteins) can reveal the 3-D structure, conformational dynamics, reaction rate, and chemical environment of proteins as well as proteins or ligands associated with NPs.¹¹⁰ Raman spectroscopy (resolution: sub-µm laterally and µm longitudinally) and X-ray photoelectron spectroscopy (XPS; detection limit: ~100 ppm, for top sample surface), techniques commonly used for elemental analysis and surface science, may find ample use in the study of intricate and adaptive relationships between NPs, the biocorona as well as host cells and organisms.^{110,111} In addition, thermogravimetric analysis (TGA; resolution: ~1 mg), an ensemble technique commonly used for determining selected characteristics of materials that exhibit either loss or gain in mass due to decomposition, oxidation, or loss of volatiles, should find use in quantifying the biocorona adsorbed on NPs of large quantities (>1 mg) as well as mass loss of NPs themselves due to biodegradation.¹³²

Limitations

The complex composition (*i.e.*, proteins, lipids, fatty acids, NOM, *etc.*) and cell/organism-dependent nature of the biocorona require extra care in their characterization. While

increases in the beta sheet content as a result of protein aggregation are often characterized by the fluorescence-based thioflavin T and Congo red assays,^{133,134} these methods remain to be validated for dealing with heterogeneous proteinous substances as is the case with the biocorona. SPR is effective in extracting the rates of binding between proteins and NPs, but its applicability to heterogeneous and complex protein samples has vet to be developed for targeting environmental biocoronae. CD spectroscopy reports changes in protein secondary structure at the ensemble level ($\sim 0.1 \text{ mg mL}^{-1}$, for 1 mm cells) and would require rigorous controls to extract such information for the protein species of interest within the entity of the biocorona. ITC is designed to measure minute thermal exchanges between small ligands and macromolecules of ~10-500 µM which, although rewarding in thermodynamic data collection, remains a challenge for characterizing highly heterogeneous biocorona (multiple species and molecular weights) and NPs (non-uniform size, shape and surface coating). With proper sample preparation and staining, TEM can resolve both the presence and thickness of the biocorona on NPs. Cryo-TEM, specifically, is suitable for examining soft biocorona weakly associated with NPs, cells or organisms as dehydration could alter or eliminate such affinity if mediated via hydrogen bonding, van der Waals forces or hydrophobic interactions. X-ray crystallography is a highly specialized. laborious and expensive technique for characterization of specific materials that form crystals in specific solvents. Scanning electron microscopy (SEM; resolution: $\sim 10 \text{ nm}$)⁹¹ (Fig. 2D, inset) equipped with elemental analysis is appropriate for imaging biocoronae associated with cells or organisms, as electrons reflected off the surfaces of the cells and organisms interfere to render three-dimensionallike topologies.

Potential

Hyperspectral imaging is a technique which has attracted increased attention recently within the field of nanoEHS (Fig. 2E). The principle of dark-field imaging entails that such a technique is suitable for examining metal and metal oxide NPs with or without the coating of biocorona, as proteinous substances alter the dielectric constant and hence SPR of the NPs in terms of amplitude and/or peak wavelength. When incorporated with a scanner along the optical axis - a newly developed feature among such systems hyperspectral imaging is suited for screening the adsorption, uptake and intracellular distribution of NPs, especially for vacuoles, cells, and tissues as well as small organisms such as bacteria, algae, protozoa, and daphnids. This sectioning feature resembles that of confocal microscopy, albeit with lower spatial resolution due to the diffraction limit along the optical axis without a pinhole. However, laborious calibrations of nano- and biomaterials to establish spectral references^{128,129} are a prerequisite of implementing data analysis for each nano-bio system under study.

UV-vis spectrophotometry is a common but often overlooked technique for quantifying protein concentration as well as protein–protein and protein–NP interactions. This method can be subsequently plugged into the Freundlich isotherm to derive multiple-layer adsorption of an adsorbate (*i.e.* biocorona) onto a rough adsorbent surface with multiple binding sites (*i.e.* NPs, cells or organisms). Based on this principle, Lin *et al.* and Bhattacharya *et al.* examined the inhibitive effects of styrene nanoplastic and CdSe/ZnS quantum dots on the photosynthetic properties of freshwater algae *Chlorella*, freshwater/saltwater *Scenedesmus* and *Chlamydomonas* sp., respectively.^{89,135}

The use of DLS for the quantification of the size and uniformity of NPs has become a standard practice in nanoEHS over the past decade. Recently, Wang et al. experimented on the use of high-throughput DLS for characterizing the binding of branched polyethyleneimine-coated silver NPs (bAgNPs) with cationic lysozyme and anionic alpha lactalbumin (aLact).¹³⁶ With 384-well plates, this technique was capable of resolving the continuously changing hydrodynamic sizes of bAgNPs exposed to the protein species with a range of concentrations over 24 h, consuming merely 6 µL of each sample volume. Furthermore, the samples were heated and then cooled at a pace of 0.52 °C min⁻¹ via a temperature control module, displaying a remarkable hysteresis in the hydrodynamic size of lysozyme-bAgNP complexation versus temperature to reveal the intricate thermostability of the nanobio system. This facile and high-throughput technique (Fig. 2F)¹²⁶ may find its use for characterising ENMs and the biocorona, allowing many different sample conditions to be monitored simultaneously and continuously at different pH values, temperatures, ionic strengths/hardnesses of freshwater and salt water.

Another high-throughput technique which we recommend for analysing the biocorona is differential scanning fluorometry (DSF), a fluorescence-based thermal shift assay which employs a hydrophobic fluorophore such as Sypro Orange and multiwell plates (Fig. 2G).¹³⁷ As a result of unfolding at increased temperatures, the hydrophobic core of a molten protein of interest becomes exposed to the external environment. Consequently, the fluorophore binds to the protein core and, upon excitation, gives rise to fluorescence that is otherwise quenched by the aqueous solution. Further increases in temperature cause the proteins to form aggregates through intermolecular hydrophobic interactions, which in turn inhibits incorporation of the fluorophore into the proteins to gradually cease fluorescence. Using this technique, Chen et al. uncovered the contrasting effects of fullerol $C_{60}(OH)_{20}$ on the thermal stability of lysozyme and immunoglobulin G (IgG) at different stoichiometric ratios and attributed those to the physicochemical properties of the proteins as well as the duality of fullerol as both a particle and a chemical.¹³⁷ Such duality originates from the limited solubility and finite size of ENMs and may have profound implications for the configuration and stability of the biocorona as well as the toxicity and fate of NPs in the environment.

The combined use of different analytical techniques can prove synergistic for examining NP-biocorona and cell/ organism-biocorona interactions. For example, Chen et al. applied the techniques SEM, inductively coupled plasma mass spectrometry (ICP-MS), and membrane potential and viability assays, as well as attenuated total reflectance (ATR)-FTIR to illustrate the adaptive interactions between freshwater algae Chlorella sp. and ZnO NPs (Fig. 2D and H), including agglomeration of algae to reduce their surface area upon NP exposure, secretion of algal exudates to cause aggregation and consequently reduced ion release and toxicity of the NPs, and formation of new bonds between algal exudates and the ZnO NPs.⁹¹ This approach may be extrapolated for assessing the interactions between other classes of aquatic organisms and metal oxide NPs, which will help establish a full database for mitigating environmental release of discharged ENMs.

Focusing on biocorona – a perspective from statistical modelling

Statistical modelling in nanoEHS

The application of statistical modelling in the nanoEHS field was made possible with the advancement of HTS and characterization technologies and instrumentation, as well as considerable amounts of data from active experimental research accumulated in the past decade. The biological activities or environmental effects of ENMs are considered dependent upon the intrinsic physicochemical properties of the ENMs. However, depending on the experimental methods, host systems' settings, and potential reactants, those properties or attributes of ENMs can make up a large matrix. It is sometimes difficult to correlate property matrices obtained from materials characterization with observed biological or environmental effects or outcomes in a sensible way and make interpretations or predictions accordingly. Statistical processing and modelling of the data allows determination of the relationships between material properties as inputs and the resulting effects as outputs in a quantitative and predictive manner (Fig. 3). Based on such principles, a statistical modelling approach, known as QSAR,^{138,139} can be constructed using parameterized quantitative property descriptors to predict biological or environmental interactions with various species from the host systems.

As an important implementation of supervised statistical learning in nanoEHS, QSAR algorithms learn from actual experimental data, construct predictive models utilizing large sets of experimental data as a training set, and choose the most relevant parameters based on a given model form and statistical criteria such as goodness-of-fit and cross-validation. Thus, the parameter coefficients can be calculated, while predictions and comparisons for the sets that are not included in the training set can be made. In this process, information on determinative ENM properties can be extracted by evaluating the weight of each parameter represented by the coefficients.¹⁴⁰ The specific field of application depends on the type of data used for training, as well as the model structure. For example, the linear solvation energy relationship (LSER) linearly relates the solvation energy parameters of a given chemical or nanomaterial to a certain activity, typically partition between two media such as adsorption. LSER is often used to find the determining properties of ENMs in



Fig. 3 Schematic of statistical modelling for nanoEHS research. Experimental data obtained from physicochemical interactions with biological or organic molecules, as well as biological effects of nanomaterials, can be collected and curated as a large database. Statistical modelling can be built to characterize the surface physicochemical properties of nanomaterials and explore how they govern the interactions with surrounding molecules. Statistical learning methods are employed for both predictive modelling and classifications based on either surface properties or biological effect profiles.

organic toxicant adsorption for environmental applications.141,142 Applying supervised statistical learning to discrete data generates classification models that can provide estimations on categorical or nominal outcomes. One of the most widely used classification algorithms, Support Vector Machine (SVM), adopts a classifier and adjusts its flexibility to reach an optimized point based on the training data, to ensure that the virtual distance from each data point of different categories to the classifier in the abstract data space reaches a maximum.^{143,144} For example, since genotoxic materials can cause cancer, statistical classification modelling on genotoxicity data sets of a given group of chemicals can be used to determine the possibility of a given material to be carcinogenic.145 SVM has been adopted for pattern recognition in discriminating potential drug chemicals based on their induced biological responses^{146,147} and should find use in describing the relationship between biocorona and the fate of cells and organisms upon ENM exposure.

Unsupervised statistical learning methods, such as clustering, have also been adopted to analyse nanoEHS data. Clustering techniques are often capable of extracting information from input data that are unlabelled or without expected outcomes. For example, hierarchical clustering was used to analyse data of protein associations with gold and silver NPs.¹¹⁶ This technique can identify latent similarities among the NPs based on their protein fingerprints and then group them hierarchically. This process does not require a known classification of the NPs, but rather creates a categorization based on their similarities in terms of protein binding. Unsupervised statistical learning could prove advantageous when dealing with heterogeneous biocoronae secreted by multiple cell or organism species exposed to one or more types of ENMs.

Statistical modelling on physicochemical interactions

Characterisation and prediction of physicochemical interactions between ENMs and various organic or biological species in the host system, especially surface adsorption or the formation of biocorona, is a major effort in the nanoEHS applications of statistical modelling. These interactions are typically governed by various fundamental physicochemical forces whose ensemble effects can usually be determined experimentally, while individual forces of different origins often cannot be distinguished.¹⁴⁸⁻¹⁵⁰ QSAR has been shown as a implementation of statistical modelling successful connecting structures of chemical species to their environmental and biological effects, including drug design, pharmacokinetics, toxicology, and medicinal chemistry,151-156 as well as analytical and physical chemical studies.^{153,157-159} To determine surface physicochemical interactions, LSER, as an implementation of QSAR, is often adopted, based on the assumption that surface binding is linearly correlated with several parameterized physical forces that likely participate in the process. For example, in a proposed biological surface adsorption index (BSAI) approach, Abraham's solvatochromic descriptors that parameterize interactions arising from lonepair electrons, hydrogen bonding, electrostatic forces, and hydrophobic forces have been used to characterize surface adsorption of environmentally relevant organic pollutant molecules to porous carbon materials, as well as metallic or metal oxide ENMs.^{140,160-162} Specifically, the binding coefficients are obtained experimentally from a set of organic molecules with varying chemical properties, then a multivariate linear regression algorithm is used to generate the coefficients for each descriptor. Since model building based on these descriptors is essentially a supervised statistical learning process, the molecules used to obtain experimental data for the training of the model need to expand a reasonably large chemical space, in order to ensure the applicability of the resulting model to the prediction of molecules that are not included in the training sets. For interactions that may involve nonlinear processes, such as crowding of biocorona on a NP surface over time or competitive binding of multiple solute species, artificial neural networks (ANN) could be an ideal method to reveal latent relationships between the predictors and the outcome, since ANN adopt nonlinear functions within one or more hidden layers between input data and output coefficients. For example, compared to linear regression, the prediction capability of surface binding of organic pesticides onto NP surfaces was improved using ANN.¹⁶²

In addition to prediction, the modelling results can also be used to evaluate the surface properties of ENMs, since the coefficients represent the likelihood of each physicochemical force to participate in the interaction.¹⁶³ Therefore, the model could potentially provide some insight into how ENM surfaces interact with biomacromolecules. The advantage of using solvatochromic parameters in modelling is that they can be easily related to physicochemical interactions and thus are ready for physical interpretations. However, for the purpose of prediction, other parameters are available, e.g., one study compared molecular connectivity-based indices to solvatochromic parameter-based LFER models and vindicated the former as of higher predictive capability.¹⁶⁴ In addition, comparing BSAI-predicted values with results from molecular dynamics (MD) simulations could help validate the molecular models used in the simulations, which in turn can be applied to prediction without experimental data.¹⁶⁵

Statistical modelling on biological effects

As aforementioned, statistical modelling often employs machine learning algorithms and its application largely depends on the data to be learned from. Similarly, QSAR can be used for the evaluation of *in vitro* biological effects of ENMs using high-throughput cell-based techniques. Similar statistical learning algorithms can be employed to study the correlation between ENM surface physicochemical properties and biological effects, including cell uptake,¹⁶⁶ cytotoxicity,^{167–169} and embryonic toxicity.¹⁷⁰ In these studies, both experimentally measured properties such as size distribution, relaxivity (for magnetic materials), zeta potential¹⁶⁷ and even quantum mechanical descriptors such as band gaps¹⁶⁸ for semiconductors and calculated molecular descriptors including molecular weight, geometric parameters, acidity/basicity and lipophilic index¹⁶⁷ were used as variables in respective statistical models. The quantum mechanical descriptors are important because they are especially relevant in determining redox reactions and ROS generation in biological systems.

By and large, the selection of these descriptors needs to reflect the roles played by their respective physicochemical properties. Typically, a large number of initial descriptors should be included, then the number can be algorithmically reduced by fitting the data. For example, Chau *et al.* initially introduced 679 descriptors, then reduced those to 367 by determining the redundancy and removing the irrelevant ones. One simple way of achieving this is through stepwise regression, which iteratively adds and removes descriptors into and from the model regression to find an optimal combination based on statistical criteria.¹⁶¹

Although the multivariate linear model is still widely used for the prediction of biological effects,^{168,171-173} due to the significant nonlinearity in biological systems more often nonlinear algorithms¹⁶⁶ are used. Logistic regression, SVM, k-nearest neighbours $(kNN)^{166}$ and partial least squares (PLS)¹⁷⁴ are usually used for classification. Association rule learning, which discovers regularities in large-scale data and generates data-driven hypothesis, was used to learn cellular responses induced by ZnO NPs.¹⁷⁵ These classification techniques are typically used when nominal or categorical outcomes are desired in describing the biological or environmental effects, such as predictions of NPs being toxic vs. non-toxic or carcinogenic vs. non-carcinogenic. For characterization of biocorona formation, supervised learning algorithms including these classification methods can categorise ENMs or biomolecules based on whether these molecules are present in the corona.

An emerging modelling approach is to use biological 'fingerprints' of a certain ENM, instead of direct physicochemical parameters, as variables for predicting how the ENM could interact with larger biological systems. Similar to the application of QSAR in determining the physicochemical interactions by using organic molecules as probes, biomacromolecules, such as proteins, are used as probes.¹¹⁶ The protein profile in a biocorona that forms on a NP surface is regarded as a set of fingerprints that describes how the surface of the NP interacts with biological components. These relative abundances from the profile are then used as predictors in a log-linear model to form a correlation with cellular association of the NPs. Subsequently, redundancy in the descriptors can be reduced by running the model through predictor selection algorithms. Closely associated with experimental data, statistical models are anticipated to be broadly applicable to the different aspects of nanoEHS, while specific applications of the models depend on the type of data. Hence, by incorporating various nano-bio or nanoenvironmental data, such as surface binding of various environmental pollutants, therapeutic chemicals or biological

macromolecules, the resulting statistical models may be used for describing or predicting their corresponding NP-ligand interactions.

Statistical learning in the big data era

Statistical learning, as a data-based modelling approach, can reveal its true potential when applied to extremely large datasets. With the development of fast and high-throughput analytical tools and accumulation of years of experimental studies, a comprehensive database for all nanoEHS-related experimental data on most existing ENM categories can be established through worldwide collaboration, for the purpose of data analysis, organization, archiving and sharing. This strategy is especially promising for building predictive statistical models by incorporating informatics tools and for predicting potential physicochemical interactions and biological effects for both medical and nanoEHS applications.

A large experimental database with multiple contributing sources requires careful curation to ensure the validity and completeness of the data.¹⁷⁶ This requires detailed documentation of experimental conditions and reaction times during the experimental studies. For example, the degree of dispersion of ENMs in the aqueous phase is usually critical for studies of physicochemical interactions, while dispersion is heavily influenced by the dispersion media, ionic strength and temperature. The comparison between data obtained under different conditions or models built excluding these factors would be considered invalid. ENM surface interactions, such as surface physisorption and biocorona formation, occur over time, while ENMs themselves - especially those of assembled nanostructures with complex internal compositions - could change over time due to agglomeration, dissolution and sedimentation. As a result, the time scales at which experiments are performed should also be included for the determination of equilibrium points and kinetics. In addition, when performing statistical learning on data collected from various sources, identification and exclusion of outliers due to batch-to-batch variations and false positive signals (from ENMs interacting with the chemicals in bioassays) may be extremely important for the generation of robust and reliable predictive models. With abundant data available in the literature, another more feasible approach is to collect and analyse the existing data in nanoEHS. For example, Xu et al. developed a toolkit termed Nanomaterial Environmental Impact data Miner (NEIMiner), which includes a function for automatically scraping the web for relevant and heterogeneous ENM environmental data, subsequently establishing complex data structures for modelling.¹⁷⁷ In addition, despite the limited amount of data available, a few working nanoEHS databases have been set up and running. For example, the NBI Knowledgebase (http://nbi.oregonstate.edu/) provides a comprehensive ENM library and a nano-bio interaction database categorised by nanomaterial properties such as core materials and surface functionalisations. Nanoinfo (nanoinfo. org) provides a central ENM safety database and modelling tools for assessing their environmental impacts. The Cancer Nanotechnology Laboratory portal (https://cananolab.nci.nih. gov/caNanoLab/) facilitates sharing of data on ENM characterisation and *in vitro* and *in vivo* assays.

Focusing on biocorona – a perspective from atomistic simulations

Proteins are usually the most abundant species within a biocorona. The biological identity of a NP is represented by convoluted properties of the NP and its bound proteins, rather than the NP itself on which the majority of basic and toxicological research has been focused to date.^{112,178,179} Such a NP-protein corona, whose formation and evolution primarily depend on the physicochemical properties of both NPs and proteins,^{140,180–182} may initiate its contact with the cell to trigger a cellular response. For instance, the proteins constituting the corona can interact with membrane receptors and initiate active cell uptake via endocytosis. The presence of protein corona can cause an altered toxic response.^{54,183} Formation of biocorona on the NP surface can also block functional ligands from interacting with targeted receptors, attenuating the designed functionality.¹⁸⁴ On the other hand, interactions between NPs and proteins in the corona can also affect the structure, dynamics, and function of the protein constituents.¹⁸⁵ The induced protein conformational changes might inactivate their native functions, and cause protein misfolding and aggregation.¹²² It has been shown that NPs can nucleate amyloid aggregates of proteins and possibly contribute to the development of protein misfolding diseases such as Alzheimer's and Huntington's diseases, type 2 diabetes, and also the dialysis-related amyloidosis.¹²² The exposure of otherwise buried protein segments can also be targeted as foreign peptides to elicit an immune response.114,186,187 Hence, from the perspectives of health and safety, it is important to uncover the structure and dynamics of the biocorona in order to establish correlations between the properties of pristine NPs and their biological and pathological functions.

Complementarity between computational modelling and experimental characterization

Previous research efforts have provided much insight into the various properties of the NP-protein corona, such as the existence and size of the biocorona,⁸ protein composition and evolution on the NP surface,^{119,120,181,188} impact of NP size and surface properties,^{53,189-192} and the corresponding cellular responses.¹⁹³⁻¹⁹⁵ Protein binding with NPs is rather dynamic and the bound proteins can be partitioned into "hard" and "soft" coronae depending on binding affinities and exchange rates with the solution. For example, despite thousands of protein types existing in the blood serum, only a few tens of proteins are found enriched in the hard corona,^{53,119,120,189-192} a major determinant for the interactions between NPs and biological systems. However, molecular details of the NP-protein corona—such as the orientations of proteins bound to the NP, protein conformational changes upon NP binding, protein-protein association and aggregation on the NP surface, and the corresponding physicochemical determinants of these complex interactions-are largely unknown due to challenges in high-resolution experimental characterization, including NP-NP agglomeration, inherent heterogeneity of the system, and the absence of solvent interactions in electron microscopy imaging. Accordingly, computational modelling can be used together with experimental studies to uncover the structural and dynamic details of complex molecular systems. By offering not only molecular insight into experimental observations but also experimentally testable hypotheses, computational modelling can bridge the gap between experimental phenomena and the underlying molecular systems of interest,¹⁹⁶ thereby making significant contributions to our understanding of the structure, dynamics and function interrelationship of the biocorona and their biological and environmental implications.

Computational methodologies for studying the structure and dynamics of the nano-bio interface

A number of computational approaches that have been originally developed to study biomolecules in computational biophysics, such as quantum mechanics simulations (OM), molecular mechanics (MM/MD), molecular docking, dissipative particle dynamics (DPD), and their combinations, 197,198 have already been utilized to study the structural and dynamic properties of the NP-protein corona at the molecular and atomic levels. Density functional theory (DFT)-based QM calculations can accurately describe the interaction energies of various configurational states of a molecular system, but are computationally expensive. As a result, QM studies are often limited to interactions of small chemical groups with a NP surface or a small nano-cluster,¹⁹⁸ which can be used to parameterize the classical mechanics interaction potentials between the NP and protein atoms for MD/MD simulations,¹⁹⁹⁻²⁰¹ or in conjunction with MD simulations of the binding proteins (i.e., hybrid QM/MM simulations) to study NP-protein complexes.197

The MM/MD simulations have been thus far the most popular approach to study the NP-protein corona.^{101,198,202-206} The traditional all-atom MD methods with explicit solvent models are capable of accurately describing molecular systems consisting of NPs and proteins under aqueous conditions and have been successfully applied to describe the binding of proteins with fullerene,^{202,203} carbon nanotubes¹⁰¹ and, more recently, metal NPs.¹⁹⁸ Given the challenges of large system size and long timescales associated with biocorona formation and evolution, it is still difficult for allatom MD simulations to reach the time and length scales required for depicting large NP-protein systems till equilibration.^{207,208} Coarse-grained (CG) MD simulations,²⁰⁹ where certain groups of atoms are simplified as single pseudo-"atoms" and effective interactions are assigned between these atoms, can be used to study large systems and reach long

time scales.^{210,211} Enhanced sampling methods such as GPUs²¹² have also been explored for CG MD simulations to study corona formation between NPs and multiple proteins. While CG simulations have been applied to study the general aspects of NP-protein interactions,^{204,209,212-216} they often have limited predictive power and lack sufficient molecular details for modelling specific molecular systems. Multiscale approaches,²¹⁷ where long time-scale simulations with CG models and short time-scale simulations with the all-atom model are coherently blended together,^{218,219} have been applied to study the effect of NP surface patterns on protein binding²⁰⁴ and characterize the structure and dynamics of biocorona formation with AgNPs.^{100,205}

Discrete molecular dynamics (DMD) is a special type of MD approach that has also been applied to study the nanobio interface.^{126,136,205,206,220-222} DMD methods use discontinuous instead of continuous potential functions to describe inter-atomic interactions,^{223,224} which improves the computational efficiency by avoiding frequent calculations of acceleration and position updates in the traditional MD approach. Both all-atom^{136,221,225} and CG^{217,222} DMD methods have been developed to model the nano-bio interface. With the high computational efficiency of DMD and application of the multiscale modelling approach, Ding et al. reported the formation of single- and multi-layer protein coronae on an AgNP surface²⁰⁵ and uncovered stretched-exponential protein binding kinetics observed experimentally.²²⁶ Qing and Carol applied CG DMD simulations to study the isotherm properties of protein adsorption to a NP using a large number of CG proteins.²²² DMD simulations have also been recently applied study the effect of NPs on protein amyloid to aggregation.126,221,227

Other methods like molecular docking and DPD have also been applied to study the possible binding between NPs and proteins. Since docking simulations usually assume a rigid protein receptor, this method cannot be used to study the conformational changes of proteins often observed upon NP binding. DPD methods usually use CG representation of the molecular system and are often used to model the mesoscopic phenomena instead of the structure and dynamics of the NP-protein corona at the atomic level. For instance, applications of DPD in the uptake of NPs with biomembranes have been recently reviewed.²²⁸

Applications of computational modelling in conjunction with biophysical and biochemical characterization have already provided much insight at the atomic and molecular levels, such as binding poses of proteins in the corona,^{100,198,205} conformational changes of proteins upon binding with NPs and nanostructures,^{101,202,207,229} and effects of the NP size, shape,²³⁰ and surface chemistry²⁰⁶ on protein binding. Due to computational limitations, most of the computational studies in the literature aimed to uncover highresolution structural and dynamic information have been focused on simulations of single proteins interacting with a single NP, which is usually not the case *in vitro* and *in vivo*. Recently, computational modelling efforts have been devoted to understanding the effects of NPs on complex biological processes involving multiple biomolecules, such as proteinprotein recognition^{231,232} and protein aggregation.^{215,227,233,234} In the following section, we will focus on recent computational and modelling studies concerning the effects of NPs on protein aggregation, where efficient modelling of multiple proteins in the presence of NPs *in silico* becomes necessary. Understanding the physicochemical determinants of NPs on protein aggregation may help design "safe" non-amyloidogenic NPs or nanomedicine with anti-amyloid properties, thereby contributing to the central goal of nanoEHS.

Complex effects of NPs on protein amyloid aggregation

Protein misfolding and amyloid aggregation are associated with a wide range of human diseases, including Alzheimer's, Parkinson's, and Huntington's diseases and type-2 diabetes.²³⁵⁻²³⁸ Despite their differences in sequence and structure, amyloidogenic proteins associated with these diverse diseases assume the same characteristic cross-β structure in their final fibrils, where misfolded proteins form extended β -sheets along the fibril axes.²³⁹⁻²⁴⁴ Many recent studies suggest that intermediate oligomers populated along the aggregation pathway, rather than the final amyloid fibrils, are the toxic species.²⁴⁵⁻²⁴⁷ Motivated by the possibility of NPs crossing the blood-brain barrier,²⁴⁸⁻²⁵¹ much effort has been devoted to understanding the effects of NPs on protein aggregation in neurodegenerative diseases, such as those of amyloid-beta peptide in Alzheimer's disease, 122, 123, 215, 233, 252-257 islet amyloid polypeptide (IAPP) in type 2 diabetes,^{126,221} and beta-2 microglobulin in dialysis-associated amyloidosis.122 Given the advancements in nanotechnology and nanomedicine, many studies have also been focused on understanding the impact of ENMs on protein aggregation. For instance, drug delivery using NPs with inhibitory effects on amyloid aggregation may have additional therapeutic effects in treating protein misfolding diseases.

Protein aggregation is a highly complex biological process depending on many physicochemical properties of both the aggregating proteins (natively folded or intrinsically disordered, thermostability and concentration) and their environments (pH, temperature and ionic strength). Usually, amyloid aggregation can be described by a nucleation process characterized by a lag time followed by a sigmoidal increase in amyloid fibrils (Fig. 4A).²⁵⁸ The nucleation step is initiated by the formation of critical oligomers of the amyloidogenic protein or peptide monomers (a natively folded protein usually undergoes partial unfolding into an aggregation-prone state²⁵⁹). With the addition of more monomers, the oligomers elongate into proto-fibrils, corresponding to a rapid increase in aggregates. Mature fibrils are formed by bundling of multiple protofibrils. Since the introduction of NPs can interfere each of these steps depending on the NP's affinity for various molecular species along the aggregation pathway, both aggregation promotion and inhibition have been reported.^{122,233,234,252,255-258} Using CG DMD simulations, it was demonstrated that differences



Fig. 4 Effects of NPs on protein aggregation. (A) Sigmoidal amyloid aggregation kinetics. A natively folded protein (brown sphere) undergoes partial unfolding or misfolding to the amyloidogenic state (cyan sphere). The nucleation process corresponds to the formation of critical oligomers (the species with the highest free energy barrier). A NP interferes with the aggregation process by interacting differentially with the molecular species along the aggregation pathway. (B) Dependence of amyloid aggregation on NP-protein interaction strength. The left inset depicts aggregation of peptides on a NP surface, while the right one shows proteins fully interacting with the NP rather than with other proteins.²²⁷ (C) Stabilization of an IAPP monomer in a helical conformation by encapsulation and binding with a PAMAM dendrimer.¹²⁶ (D) Formation of "off-pathway" oligomers from polyphenol nano-clusters.²⁶⁰ (E) Disruption of amyloid fibrils by graphene.²³⁴ (F) Competitive binding of natural amphiphiles – *i.e.*, cellulose (red), peptide (blue), palmitic acid (cyan) – with a graphene oxide nanosheet (yellow).²²⁰ (G) The number density of molecules, *n*, bound to the nanosheet as a function of time.²²⁰

in NP-protein binding affinity and their relative concentration contributed to the complex NP effect on amyloid aggregation (Fig. 4B).²²⁷ Atomistic MD simulations have also been applied to explore the inhibition of amyloid aggregation by targeting various molecular species along the aggregation pathway (Fig. 4A), including the monomers (Fig. 4C),¹²⁶ oligomers (Fig. 4D),²⁶⁰ and also the amyloid fibrils (Fig. 4E).²³⁴

Atomistic DMD simulations have recently been applied to elucidate the competitive binding of natural amphiphiles including celluloses, peptides, and lipids with graphene oxide nanosheets (GRO),²²⁰ where lipids and peptides displayed stronger binding to GRO than celluloses. Given the differen-

tial population of natural amphiphiles in the environments (*e.g.*, cellulose populates the algal exudates the most, followed by peptides and lipids), a Vroman-like competitive binding phenomenon was observed, where the most abundant celluloses bound to GRO first but were later replaced by stronger binding lipids with lower abundance (Fig. 4F and G).

Despite the capacity of computer simulations in revealing the inner workings of the biocorona that is often beyond the resolution of statistical modelling, including the composition, crowding and aggregation of the biocorona in response to time, temperature, pH, salt strength, heterogeneity and concentration of natural amphiphiles, pollutants and exudates, current research on nanoEHS lacks such a viable computational element. As the structure, dynamics and function of the biocorona entail rich toxicological and pathological implications, the value of computer simulations will likely be increasingly acknowledged and utilized for advancing nanoEHS research in the new era.

Outlook

To ensure its sustainability, research on nanoEHS should be purposefully aligned with ENM synthesis and move beyond simple materials to sophisticated and smart ENMs used in real-world applications.^{261,262} For example, there has been a vast pool of polymeric and polymeric-condensed hybrid nanocarriers and nanoconstructs developed for gene and drug delivery.¹¹⁰ Such nanostructures are often grafted with polyethylene glycol (PEG), poly(alkyl cyanoacrylate) or poly(lactic acid) to avoid protein fouling and prolong their blood circulation. However, PEG is not biodegradable and has been found prone to the adsorption of apolipoproteins A-1 and E as well as complement proteins that play important roles in immunogenicity.²⁶³⁻²⁶⁶ Furthermore, these nanostructures have rarely been examined beyond their in vitro and in vivo toxicities determined by fluorescence and histology assays. Indeed, the efficacy of drug delivery has not improved over the past two decades despite global investment and active research,²⁶⁷ partly because of a lack of understanding of nanobiomolecular interactions and their implications for the whole organism level. In this regard, collaboration between nanomedicine and nanoEHS could prove cost-effective and mutually beneficial.

Antibacterial applications of ENMs are another area which could benefit from improved understanding of nano-bio interactions.95,268-275 Current ENM-based antibacterial applications usually rely on mechanisms acting through toxic ions, ROS, oxidative stress and membrane damage elicited by NPs (e.g. Ag, CuO and ZnO NPs, carbon nanotubes and graphene derivatives) bound to the bacterial cell wall or taken up by bacterial cells.^{21,22} Further development in this arena could exploit interactions of novel ENMs with bacterial EPS as well as inhibition of functional amyloids associated with bacterial genera such as certain E. coli and Pseudomonas strains.²⁷⁶⁻²⁷⁹ In addition, with highly active research on green chemistry and biosynthesis of antibacterial ENMs using extracts and extracellular substances from various organisms,^{280,281} it is increasingly important to focus on the biocorona formed during NP synthesis to avoid introducing potentially cytotoxic biomolecules. This area is potentially rewarding scientifically, with far-reaching implications in medicine, food science, agriculture, wastewater treatment, and environmental remediation and engineering.

With the rapid improvement of computational power, there is increasing potential in using computer simulations and statistical modelling for research on nanoEHS. Both approaches can take into consideration long-range electrostatics (pH, water hardness, and charges of NPs and natural amphiphiles), short-range van der Waals forces, hydrophobic interactions, hydrogen bonding and $\pi - \pi$ stacking, as well as the aspects of protein/lipid concentration, crowding, and temperature that varies from season to season and region to region.²⁸² Recent development in molecular modelling of the nano-bio interface has significantly expanded the boundary of single NP-protein complexes to larger, more sophisticated, and more realistic systems. As large experimental data are being generated at an increasing rate, especially with the incorporation of industrial automation into research laboratories, advanced statistical learning algorithms are needed to extract valuable information to understand the nature of the interactions and make further assessments and predictions. In addition, applications of methodologies adopted from proteomics and lipidomics in nanoEHS often generate a large volume of mass spectral data, which can be viewed as the biomacromolecular fingerprints of respective ENMs. Towards that end, advanced informatics algorithms with big data capabilities are essential for the advancement of nanoEHS.

We envisage that a broad range of nanotechnology-based environmental applications would benefit from the integration of the tools discussed above. The development of simulation methodologies will allow systematic design and manufacturing of nano-enabled products from basic nanobio interaction principles, e.g. membranes for water and air filtration could be designed specifically based on the affinity between the membrane materials and intended pollutants. The established QSAR profiles of ENMs could be used as guidelines for the design of catalytic treatment based on redox-active NPs or nanostructures. We believe that highthroughput and predictive toxicological methods can significantly contribute to elucidating adverse outcome pathways of ENMs that are gaining increasing scrutiny²⁸³ and are proposed to be incorporated into risk assessment of ENMs.²⁸⁴ Recently, the surface affinity and dissolution rate were identified as critical functional assay parameters for characterizing ENM behavior.²⁸⁵ We anticipate that biocorona characterization with experimental methods, modelling and simulations would greatly advance nanoEHS research and assist ENM risk assessment. Taking advantage of predictive toxicological approaches and high-throughput screening platforms, we anticipate integration of safety assessment into mainstream nanotechnology research to support safe manufacturing. Knowledge generated through such endeavours will aid not only the first generation of ENMs but also the new generation of advanced nanostructures and nanosystems.

With the fusion of expertise from structural biology, biophysics, proteomics, analytical and polymer chemistry, computational science and technology, pharmaceutical sciences and nanomedicine, nanoEHS is poised for its transition from a nascent field centred on nanotoxicology to a more mature, multidisciplinary and broadly defined area of research centred on real-world applications and relevance in the coming decade. Towards that end, understanding biocorona may prove pivotal for the design of smart ENMs with minimal ecological footprint.

Competing financial interests

The authors declare no conflicting financial interests.

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